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Habitat Associations and Landscape Ecol
OSMP Studies 4257

Study



Berry, Mark E.

HABITAT ASSOCIATIONS AND LANDSCAPE ECOLOGY
OF BREEDING BIRDS IN FOOTHILLS SHRUB IN THE NORTHERN
COLORADO FRONT RANGE

Final Report to:

City of Boulder, Dept. of Open Space / Real Estate and Division of Mt. Parks

MARK E. BERRY

University of Colorado

Department of Environmental, Population, and Organismic Biology

Boulder, CO 80309-0334

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ABSTRACT

Foothills shrub in the northern Colorado Front Range is a unique and understudied habitat type. I investigated habitat associations of breeding songbirds in patches of foothills shrub in relation to measures of vegetation structure, vegetation composition, landscape context, and proximity of recreational trails. Distributions of breeding birds were quantified using 50 m fixed-distance point counts at 84 study plots. Six total counts were conducted at each plot in May and June of 1996 and 1997, and species were categorized as breeding season residents (present in three or more counts) or non-residents at each plot. I quantified local vegetation structure and composition using a point-centered quarter technique and visual estimation of cover, and I determined the landscape context of plots by sampling 80 0.1 ha plots on eight 500 m radial transects, using aerial orthophotographs to classify habitat types. I used multiple logistic regression to analyze associations between breeding residence of each of the eight most common breeding bird species and 16 habitat or landscape variables. I recorded a total of 59 bird species during point counts, but only eight species were resident on more than 12 plots. Spotted towhees (*Pipilo maculatus*) were resident on 98% of plots and demonstrated no significant habitat associations. Brown-headed cowbirds (*Molothrus ater*) were associated with wild plum (*Prunus americana*). Green-tailed towhees (*Pipilo chlorurus*) were negatively associated with shrub cover, grass cover, and five of the most common shrub species. Virginia's warblers (*Vermivora virginiae*) were negatively associated with shrub cover, tree cover, chokecherry (*Prunus americana*), shrub distance, and height heterogeneity. Yellow-breasted chats (*Icteria virens*) and Broad-tailed hummingbirds (*Selasphorus platycercus*) were negatively associated with horizontal heterogeneity. Neither Lazuli buntings (*Passerina amoena*) nor Blue-gray gnatcatchers (*Polioptila caerulea*) were significantly associated with any individual variables. No species showed any significant association with landscape context variables measuring the landscape occurrence of shrub, grassland, and coniferous forest habitat, suggesting that shrub patches with appropriate local characteristics were used regardless of their natural landscape setting in this system. There were no significant associations between breeding residence and proximity of recreational trails, suggesting trail impacts have not been pervasive enough to affect species distributions.

Avian Habitat Associations

Habitat selection by individual species underlies observed patterns in communities (Cody, 1985). Many factors can affect avian community composition and avian diversity. Patterns of habitat use may be determined by current responses to habitat characteristics, historical factors, landscape characteristics, and human disturbance (Rosenzweig, 1995).

The bird species using a habitat will be only partially determined by features of the habitat itself. MacArthur and MacArthur (1961) demonstrated a linear relationship between the vertical complexity of vegetation and the diversity of bird species occurring within a habitat. Roth (1976) found that horizontal heterogeneity, or patchiness, of vegetation also was associated with avian diversity. Thus, on a regional scale, vertical and horizontal structural features of habitats are important determinants of species occurrence and of avian diversity (MacArthur & MacArthur, 1961; Roth, 1976; Willson, 1974). However, on a local scale, species composition of vegetation may be more important than structural variables in affecting avian community structure within a particular habitat type (Rotenberry, 1985; Tomoff, 1974).

Patches of otherwise similar habitats that occur in different landscape settings may be used by different bird species. Island biogeography theory (MacArthur & Wilson, 1967) has contributed to our understanding of avian communities in natural or artificial habitat islands. Small and/or highly isolated islands typically support fewer species than large islands and/or islands near a mainland area. The species that are absent from small habitat islands tend to be species with low population densities, high levels of population fluctuation, low dispersal ability, or specific habitat requirements (e.g., Bolger et al., 1991; Møller, 1987; Nilsson, 1986). In fragments of natural forests, avian richness is typically reduced (Ambuel & Temple, 1983; Howe, 1984; Lynch & Whigham, 1984; MacClintock et al., 1977; Robbins et al., 1989), and nest predation and brood parasitism may be increased (e.g., Wilcove, 1985; Yahner, 1988). The habitat context of fragments may be an important factor affecting either the habitat quality or the tendency of birds to disperse into the patch (Harris, 1984; Rosenzweig, 1995). As opposed to fragments of previously continuous habitat, the avian diversity of naturally patchy habitats may be high (Bock, 1997).

The activities of humans may alter patterns of avian habitat usage. Human disturbance has been shown to affect behavior, reproduction, and habitat

properties not currently managed as recreational parks, and recreational trails typically were not visible from study points. I recorded all study sites greater than 200 m from trails as having a distance of 200 m to trails.

Habitat Sampling

I sampled vegetation at each study plot between 1 July and 1 September 1996. Habitat characteristics were measured by establishing 16 sampling points located 10 m and 20 m from the plot center on each of eight cardinal compass directions, and using the point-centered quarter technique (Cottam & Curtis, 1956) at each vegetation sampling point. I measured the distance to the nearest woody shrub at least 50 cm tall in each of four quarters, identified the contacted shrub, and measured its height. For each plot, I calculated the average distance to the nearest shrub (shrub distance), the coefficient of variation of shrub distance (horizontal heterogeneity; Roth, 1976), the average shrub height (shrub height), and the coefficient of variation of shrub height (height heterogeneity). In addition, the percent canopy cover of shrubs (shrub cover), grass and herbaceous vegetation (grass cover), and trees (tree cover) within 50 m of the plot center were visually estimated.

Landscape sampling

I utilized aerial orthophotographs to quantify the habitat context within a distance of 500 m of each study point. Most orthophotographs were at a 1:2400 scale, but in some cases the only photos available were 1:4800 or 1:6000 scale. I used transect sampling on the aerial photographs, by placing eight transects (500 m ground distance) at each study point on each of the cardinal directions, and then sampling for habitat at 10 points on each transect, separated by 50 m. At each habitat sampling point, a 0.1 ha circle was examined on the photographs and the habitat was categorized into one of the following types: shrub, conifer, grassland, riparian, bare ground, open water, or developed. Only the habitat type covering the greatest area within each circle was recorded. For each study plot, I recorded the percent occurrence of each habitat type.

RESULTS

Habitat Characteristics

Thirty-two shrub taxa occurred on the 84 plots, including 27 species and five genera not identified to species (Table 1). Over 50% of shrubs contacted during point-quarter sampling were of only two species: skunkbrush and mountain mahogany, while five shrubs (adding chokecherry, wild plum, and hawthorn) accounted for nearly 80% of the observations. The remaining 27 shrubs occurred on an average of only 9 percent of plots.

The study plots were highly variable structurally. Average distance to the nearest shrub from vegetation sampling points ranged from 0.03 m to 22.37 m, while average shrub height ranged from 0.78 m to 4.04 m (Table 2). Cover of shrubs within 50 m of plot centers ranged from 15 to 89%, with a mean of 51.5%. Grass cover averaged 43.7%, while tree cover averaged 9.5%.

Grassland, shrub, and coniferous forest were the predominant landscape cover types surrounding the 84 study points (Table 3). Riparian forest, bare ground, open water (reservoirs), and developed land were substantially less common.

Resident birds

I recorded 59 bird species on study plots during six point counts over two years. (Table 4; see Appendix B for scientific names and Appendix C for individual plot occurrences) Thirty-four species occurred at least three times on at least one plot, and were considered breeding season residents on at least one plot. Only eight species were resident on more than 12 plots.

Multiple logistic regression revealed significant relationships between probability of breeding residence and at least one habitat variable for five of the eight most common species (Table 5). Spotted towhees were resident on 82 of 84 plots, and no significant habitat associations were revealed by multiple logistic regression. Brown-headed cowbirds were considered resident on 43 plots, and were significantly associated with frequency of wild plum. Green-tailed towhees were significantly negatively associated with seven habitat variables: shrub cover within 50 m, grass cover within 50 m, skunkbrush, mountain mahogany, chokecherry, wild plum, and hawthorn. Virginia's warblers were significantly

this species is apparently able to utilize (Andrews & Righter, 1992; Sedgwick, 1987; Shugart & James, 1973) .

Brown-headed cowbirds also were common, and revealed only one significant habitat association, with wild plum. Wild plum was most common in moist shrub patches which were frequently near forest or grassland edges. Cowbirds apparently prefer edge habitats to extensive uniform areas, as indicated by patterns of nest parasitism (Lowther, 1993; Chace, 1995). In southwestern Colorado Gambel oak habitats, cowbirds were most common in pure oak and mixed pinyon pine (*Pinus edulis*)-juniper (*Juniperus osteosperma*)-Ponderosa pine-oak associations (E. Stone, unpublished data).

Green-tailed towhees showed an interesting pattern, with negative associations with shrub and grass cover and all five of the most common shrub species, suggesting a preference for heterogeneous habitats not dominated by a single shrub species. Other studies of Green-tailed towhee habitat selection have yielded diverse results, including associations with mountain mahogany, skunkbrush, bitterbrush (*Purshia tridentata*), currant (*Ribes* sp.), and Rocky Mountain juniper (*Juniperus scopulorum*) in northern Colorado (Knopf et al., 1990); with open shrubby areas of high shrub species richness in northwest Colorado pinyon-juniper woodlands (Sedgwick, 1987); with rabbitbrush (*Chrysothamnus viscidiflorus*) and bitterbrush in the northwestern Great Basin (Wiens & Rotenberry, 1981); and with oak and oak-serviceberry (*Amelanchier* sp.) in southwestern Colorado (E. Stone, unpublished data).

Virginia's warblers were negatively associated with shrub cover, tree cover, chokecherry, average shrub distance, and height heterogeneity. This species seemed to prefer areas with many small, dense patches of shrubs, often with skunkbrush and scattered Rocky Mountain juniper or ponderosa pine. Virginia's warblers utilize a variety of habitats in Colorado, including Gambel oak, mountain mahogany, riparian thickets, ponderosa pine forests, and pinyon-juniper woodlands (Andrews & Righter, 1992). In northwestern Colorado pinyon-juniper woodlands, Sedgwick (1987) found Virginia's warblers in Gambel oak chaparral or draws; while in southwestern Colorado, this species was most common in oak-serviceberry, pinyon-juniper-oak, and pinyon-juniper-ponderosa-oak associations (E. Stone, unpublished data).

Lazuli buntings showed no significant habitat associations. Lazuli buntings occur in a variety of shrub habitats both in Colorado and throughout

Landscape Ecology and Conservation

No significant relationships were evident between measures of landscape context and occurrence of common breeding bird species. Patch size was not definable in this study because of the highly variable dispersion patterns of individual shrubs. Any attempt at defining patch size would have required an arbitrary selection of scale. A scale selected by a human investigator would not necessarily have been relevant to breeding songbirds (Levin, 1992; Maurer, 1985; Morris, 1987).

By contrast, fragments of deciduous forest in the eastern U. S. have typically had reduced avian species richness caused by the loss of "area-sensitive" species (Ambuel & Temple, 1983; Howe, 1984; Lynch & Whigham, 1984; MacClintock et al., 1977; Robbins et al., 1989). Herkert (1994) and Vickery et al. (1994) found that many grassland birds in Illinois and Maine were dependent on large patches of habitat. Bock et al. (In press) found that some grassland birds were unlikely to occur near suburban development at the edge of grassland patches. In western riparian systems, Gutzwiller and Anderson (1987a) and Saab (1996) found that landscape variables, including patch size and measures of landscape context, were important to species richness and the distribution of some individual species. In each case, surrounding landscapes sometimes included human-altered areas.

In addition to human-caused fragmentation, island biogeography theory also would predict small natural habitat islands to have reduced species richness relative to large islands, due to lower probabilities of colonization and higher probabilities of extinction (MacArthur & Wilson, 1967). However, if dispersal to habitat patches is frequent enough that populations are not independent (or individual patch population dynamics are not at equilibrium), metapopulation models may be more appropriate predictors of community composition than island biogeography theory (Gutzwiller & Anderson, 1987b; Pulliam, 1988; Pulliam & Danielson, 1991).

In riparian woodlands in Arizona, Strong and Bock (1992) found that patch size was not a good predictor of avian density or richness, perhaps because of the natural patchiness of these habitats. Naturally small patches may provide important resources to species with populations (or individuals) that do not exclusively utilize the individual patch.

nesting in grasslands or ponderosa pine. Further research on the potential impacts of recreational trails to birds and other wildlife in foothills shrub habitats should be carried out before additional trails are planned.

Brown-headed cowbirds occurred in more point counts in foothills shrub study plots than any species other than the Spotted towhee. The prevalence of this obligate brood parasite in the landscape strongly suggests that cowbirds may be having an impact on other breeding birds in foothills shrub. Most of the common breeding species in foothills shrub are known cowbird hosts (Chace & Cruz, 1996). Further research in foothills shrub patches should examine the effects of cowbird brood parasitism on the breeding biology of other songbird species.

Northern Colorado Front Range foothills shrub is a breeding bird habitat with substantial conservation value. Shrub patches support an extremely high density of breeding birds in comparison with other foothills habitats (Steven Jones, unpublished data). Several breeding species of foothills shrub patches are shrub specialists, including both the abundant Spotted towhees and locally less common Green-tailed towhees, Virginia's warblers, Lazuli buntings, Yellow-breasted chats, and Blue-gray gnatcatchers. While some of these species also occur in other habitat types, foothills shrub is clearly an important local habitat for each of the above species. Foothills shrub may also be important to locally rare species such as Gray catbirds (Andrews, 1979), Scrub jays, Chestnut-sided warblers, and Bushtits. Foothills shrub is a relatively rare habitat type in the northern Front Range foothills, especially in the vicinity of Boulder, and therefore may deserve a higher conservation priority than habitats with greater geographic extent, such as ponderosa pine forest. Foothills habitats on private land are vulnerable to development for suburban housing. Protection of foothills shrub patches that support a variety of breeding bird species should be a local conservation priority.

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Table 1. Shrub species recorded during habitat sampling of 84 plots in foothills shrub habitat in the northern Colorado Front Range foothills.

| Species* | Common Name | Contacts:** | | Plots:*** | |
|--|-------------------|-------------|---------|-----------|---------|
| | | Number | Percent | Number | Percent |
| <i>Rhus aromatica</i> | Skunkbrush | 1437 | 26.73 | 49 | 58.33 |
| <i>Cercocarpus montanus</i> | Mountain mahogany | 1304 | 24.26 | 38 | 45.24 |
| <i>Prunus virginiana</i> | Chokecherry | 633 | 11.77 | 58 | 69.05 |
| <i>Prunus americana</i> | Wild plum | 497 | 9.24 | 31 | 36.90 |
| <i>Crataegus</i> sp. (<i>erythropoda/macracantha</i>) | Hawthorn | 421 | 7.83 | 29 | 34.52 |
| <i>Symphoricarpos oreophilus</i> | Snowberry | 178 | 3.31 | 15 | 17.86 |
| <i>Symphoricarpos occidentalis</i> | Snowberry | 176 | 3.27 | 27 | 32.14 |
| <i>Physocarpus monogynus</i> | Small ninebark | 115 | 2.14 | 16 | 19.05 |
| <i>Ribes cereum</i> | Wax currant | 96 | 1.79 | 36 | 42.86 |
| <i>Rhus glabra</i> | Smooth sumac | 92 | 1.71 | 6 | 7.14 |
| <i>Acer glabrum</i> | Mountain maple | 90 | 1.67 | 10 | 11.90 |
| <i>Rubus deliciosus</i> | Boulder raspberry | 72 | 1.34 | 17 | 20.24 |
| <i>Rosa</i> sp. (<i>arkansana/acicularis</i>) | Rose | 64 | 1.19 | 21 | 25.00 |
| <i>Amelanchier</i> sp. | Serviceberry | 36 | 0.67 | 5 | 5.95 |
| <i>Celtis reticulata</i> | Netleaf hackberry | 34 | 0.63 | 7 | 8.33 |
| <i>Toxicodendron rydbergii</i> | Poison ivy | 32 | 0.60 | 8 | 9.52 |
| <i>Clematis ligustifolia</i> | Virgin's bower | 19 | 0.35 | 5 | 5.95 |
| <i>Chrysothamnus nauseosus</i> | Rabbitbrush | 15 | 0.28 | 6 | 7.14 |
| <i>Lonicera</i> sp. (<i>morrowi/tatarica</i>) | Honeysuckle | 10 | 0.19 | 2 | 2.38 |
| <i>Ceanothus fendleri</i> | New Jersey tea | 9 | 0.17 | 5 | 5.95 |
| <i>Viburnum lantana</i> | Wayfaring-tree | 8 | 0.15 | 1 | 1.19 |
| <i>Artemisia cana</i> | Wild sagebrush | 7 | 0.13 | 2 | 2.38 |
| <i>Rosa woodsii</i> | Woods rose | 6 | 0.11 | 4 | 4.76 |
| <i>Salix</i> sp. | Willow | 5 | 0.09 | 2 | 2.38 |
| <i>Rhamnus cathartica</i> | Buckthorn | 5 | 0.09 | 1 | 1.19 |
| <i>Rubus idaeus</i> | Wild raspberry | 4 | 0.07 | 1 | 1.19 |
| <i>Mahonia repens</i> | Oregon grape | 3 | 0.06 | 2 | 2.38 |
| <i>Colutea arborescens</i> | Bladder senna | 3 | 0.06 | 1 | 1.19 |
| <i>Ligustrum vulgare</i> | Privet | 2 | 0.04 | 1 | 1.19 |
| <i>Amorpha fruticosa</i> | Leadplant | 1 | 0.02 | 1 | 1.19 |
| <i>Ribes americanum</i> | Black currant | 1 | 0.02 | 1 | 1.19 |
| <i>Ribes inerme</i> | Common gooseberry | 1 | 0.02 | 1 | 1.19 |

* Species names follow Carter (1988)

** Maximum possible contacts = 5376 (84 plots X 64 contacts each)

*** Maximum possible plots = 84

Table 4. Detection of birds during six point counts on 84 plots in foothills shrub habitat in the northern Colorado Front Range foothills.

| Common Name* | Code | Total Number of Plots where: | | Total Number of Counts where |
|--------------------------|------|---------------------------------|----------|---------------------------------|
| | | Resident** | Detected | Detected*** |
| Spotted towhee | SPTO | 82 | 84 | 460 |
| Brown-headed cowbird | BHCO | 43 | 73 | 214 |
| Green-tailed towhee | GTTO | 33 | 59 | 175 |
| Virginia's warbler | VIWA | 32 | 54 | 148 |
| Lazuli bunting | LABU | 29 | 68 | 178 |
| Yellow-breasted chat | YBCH | 28 | 53 | 160 |
| Broad-tailed hummingbird | BTHU | 22 | 49 | 139 |
| Blue-gray gnatcatcher | BGGN | 22 | 38 | 108 |
| Black-billed magpie | BBMA | 12 | 33 | 74 |
| Western meadowlark | WEME | 12 | 24 | 72 |
| American robin | AMRO | 10 | 27 | 55 |
| Black-headed grosbeak | BHGR | 10 | 23 | 59 |
| Western tanager | WETA | 7 | 16 | 35 |
| Vesper sparrow | VESP | 7 | 14 | 33 |
| Dusky flycatcher | DUFL | 5 | 31 | 46 |
| Chipping sparrow | CHSP | 5 | 19 | 35 |
| House wren | HOWR | 5 | 19 | 35 |
| Macgillivray's warbler | MAWA | 4 | 18 | 28 |
| Warbling vireo | WAVI | 4 | 10 | 24 |
| Rock wren | ROWR | 4 | 8 | 22 |
| Gray catbird | GRCA | 3 | 11 | 22 |
| Lesser goldfinch | LEGO | 2 | 18 | 27 |
| American goldfinch | AMGO | 2 | 16 | 22 |
| Plumbeous vireo | PLVI | 2 | 11 | 20 |
| Lark sparrow | LASP | 2 | 5 | 13 |
| Mourning dove | MODO | 1 | 13 | 18 |
| Western wood pewee | WWPE | 1 | 13 | 18 |
| Pine siskin | PISI | 1 | 11 | 18 |
| Mountain chickadee | MTCH | 1 | 10 | 17 |
| Scrub jay | SCJA | 1 | 6 | 9 |
| Yellow warbler | YEWA | 1 | 6 | 9 |
| Indigo bunting | INBU | 1 | 3 | 5 |
| Song sparrow | SOSP | 1 | 2 | 5 |
| Chestnut-sided warbler | CSWA | 1 | 1 | 4 |
| Steller's jay | STJA | 0 | 13 | 16 |
| Black-capped chickadee | BCCH | 0 | 13 | 15 |
| Northern oriole | NOOR | 0 | 11 | 14 |
| House finch | HOFI | 0 | 11 | 11 |
| Brewer's sparrow | BRSP | 0 | 6 | 6 |
| Northern mockingbird | NOMO | 0 | 5 | 6 |

Table 5. Results of multiple logistic regression analyses between breeding residence of eight songbird species and 16 habitat variables. Only statistically significant individual variable associations are listed.

| | Whole Model | | Significant Associations ($p < 0.05$) | | | |
|------|----------------|----------|---|----------|----------------|---------|
| | X ² | p | Variable | Estimate | X ² | p |
| SPTO | 18.90 | 0.27 | none | | | |
| BHCO | 53.26 | < 0.0001 | Wild plum | 0.177 | 5.64 | < 0.05 |
| GTTO | 47.47 | < 0.0001 | Shrub cover | -0.173 | 12.04 | < 0.001 |
| | | | Grass cover | -0.115 | 7.67 | < 0.01 |
| | | | Skunkbrush | -0.159 | 7.42 | < 0.01 |
| | | | Mountain mahogany | -0.109 | 4.49 | < 0.05 |
| | | | Chokecherry | -0.244 | 7.31 | < 0.01 |
| | | | Wild plum | -0.186 | 8.70 | < 0.005 |
| | | | Hawthorn | -0.199 | 5.65 | < 0.05 |
| VIWA | 58.45 | < 0.0001 | Shrub cover | -0.196 | 11.45 | < 0.001 |
| | | | Tree cover | -0.178 | 7.07 | < 0.01 |
| | | | Chokecherry | -0.148 | 4.02 | < 0.05 |
| | | | Avg. distance | -0.636 | 5.11 | < 0.05 |
| | | | Height hetero. | -0.090 | 4.55 | < 0.05 |
| LABU | 26.04 | 0.05 | none | | | |
| YBCH | 39.68 | < 0.001 | Horiz. hetero. | -0.026 | 7.85 | < 0.01 |
| BTHU | 44.50 | < 0.0005 | Horiz. hetero. | -0.040 | 6.42 | < 0.05 |
| BGGN | 47.85 | < 0.0001 | none | | | |

Appendix A. Continued...

| Plot | Latitude (dec. deg.) | Longitude (dec. deg.) | Elevation (m) |
|------|----------------------|-----------------------|---------------|
| MS 1 | 40.02512014 | 105.29590765 | 1710 |
| MS 2 | 40.02792675 | 105.29813346 | 1751 |
| MS 3 | 40.03098391 | 105.29706134 | 1769 |
| SM 1 | 39.94182362 | 105.26218798 | 1727 |
| SM 2 | 39.94414047 | 105.27306637 | 1816 |
| SM 3 | 39.94302426 | 105.27996989 | 1894 |
| SM 4 | 39.93867145 | 105.28131100 | 1859 |
| SM 5 | 39.95227476 | 105.28002212 | 1899 |
| SM 6 | 39.9519649 | 105.27455746 | 1847 |
| SM 7 | 39.94940001 | 105.26987894 | 1795 |
| SM 8 | 39.95089376 | 105.26548456 | 1759 |
| WL 1 | 40.06060937 | 105.29781331 | 1825 |
| WL 2 | 40.05732942 | 105.29800462 | 1808 |
| WL 3 | 40.05037937 | 105.29867928 | 1796 |
| AP 1 | 39.7143415 | 105.21772313 | 2003 |
| AP 2 | 39.71250916 | 105.22098808 | 2028 |
| AP 3 | 39.71381201 | 105.22366590 | 2095 |
| AP 4 | 39.71614719 | 105.22228605 | 2088 |
| AP 5 | 39.71986063 | 105.22705386 | 2135 |
| AP 6 | 39.72223707 | 105.22459440 | 2054 |
| GP 1 | 39.84104197 | 105.27810740 | 2028 |
| GP 2 | 39.8332512 | 105.26808463 | 1906 |
| GP 3 | 39.8259853 | 105.26569430 | 1883 |
| HP 1 | 39.69712884 | 105.20233518 | 1957 |
| MW 1 | 39.69711299 | 105.21336220 | 2012 |
| MW 2 | 39.69817602 | 105.21821511 | 2030 |
| MW 3 | 39.68960757 | 105.20724048 | 1928 |
| MW 4 | 39.68340974 | 105.20984966 | 1958 |
| MW 5 | 39.68471226 | 105.21846873 | 2058 |
| MW 6 | 39.68671485 | 105.21678115 | 2039 |
| MW 7 | 39.68905759 | 105.21556948 | 2080 |
| RE 1 | 39.89451715 | 105.26973430 | 1969 |
| RE 2 | 39.8894159 | 105.25574487 | 1896 |
| RE 3 | 39.88705963 | 105.25813258 | 1911 |
| GL 1 | 39.75989693 | 105.24755768 | 1943 |
| GL 2 | 39.76477742 | 105.24800869 | 1990 |
| GL 3 | 39.76514436 | 105.25374604 | 2130 |
| GL 4 | 39.76734686 | 105.25068477 | 2039 |
| GL 5 | 39.76287748 | 105.24369619 | 1899 |
| WR 1 | 39.81475949 | 105.26523422 | 1940 |
| WR 2 | 39.81209117 | 105.26724882 | 2000 |
| WR 3 | 39.80835399 | 105.26678746 | 2075 |

Appendix B. Continued...

| Common Name | Code | Genus species |
|-------------------------|------|-------------------------------|
| Red-winged Blackbird | RWBL | <i>Agelaius phoeniceus</i> |
| Pygmy Nuthatch | PYNU | <i>Sitta pygmaea</i> |
| Common Grackle | COGR | <i>Quiscalus quiscula</i> |
| Wilson's Warbler | WTWA | <i>Wilsonia pusilla</i> |
| Yellow-rumped warbler | YRWA | <i>Dendroica coronata</i> |
| Grasshopper Sparrow | GRSP | <i>Ammodramus savannarum</i> |
| Cedar Waxwing | CEWA | <i>Bombycilla cedrorum</i> |
| Townsend's Solitaire | TOSO | <i>Myadestes townsendi</i> |
| White-crowned Sparrow | WCSP | <i>Zonotrichia leucophrys</i> |
| Bushtit | BUSH | <i>Psaltriparus minimus</i> |
| Blue Grosbeak | BLGR | <i>Guiraca caerulea</i> |
| Brewer's Blackbird | BRBL | <i>Euphagus cyanocephalus</i> |
| Canyon Wren | CAWR | <i>Catherpes mexicanus</i> |
| Dark-eyed Junco | DEJU | <i>Junco hyemalis</i> |
| European Starling | EUST | <i>Sturnus vulgaris</i> |
| House Sparrow | HOSP | <i>Passer domesticus</i> |
| White-breasted Nuthatch | WBNU | <i>Sitta carolinensis</i> |

Appendix C. Continued...

| Plot | SPTO | BHCO | GTTO | VIWA | LABU | YBCH | BTHU | BGGN | BBMA | WEME | AMRO | BHGR |
|------|------|------|------|------|------|------|------|------|------|------|------|------|
| SM2 | 6 | 5 | 3 | 3 | 4 | 6 | 6 | 3 | 1 | 0 | 1 | 4 |
| SM3 | 5 | 3 | 0 | 1 | 1 | 1 | 4 | 0 | 0 | 0 | 1 | 6 |
| SM4 | 6 | 0 | 4 | 1 | 5 | 5 | 6 | 2 | 0 | 0 | 1 | 4 |
| SM5 | 3 | 3 | 0 | 6 | 0 | 1 | 2 | 0 | 0 | 0 | 1 | 0 |
| SM6 | 6 | 3 | 0 | 2 | 5 | 1 | 3 | 0 | 0 | 0 | 4 | 1 |
| SM7 | 6 | 5 | 2 | 1 | 3 | 5 | 1 | 3 | 0 | 1 | 0 | 0 |
| SM8 | 5 | 2 | 3 | 0 | 2 | 1 | 0 | 3 | 1 | 3 | 0 | 0 |
| WL1 | 4 | 0 | 3 | 0 | 4 | 0 | 0 | 1 | 3 | 0 | 1 | 0 |
| WL2 | 6 | 3 | 0 | 0 | 4 | 2 | 1 | 3 | 4 | 2 | 0 | 0 |
| WL3 | 6 | 3 | 1 | 0 | 5 | 2 | 1 | 2 | 2 | 2 | 0 | 0 |
| AP1 | 6 | 1 | 6 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 2 | 0 |
| AP2 | 4 | 1 | 1 | 0 | 3 | 0 | 1 | 1 | 1 | 0 | 0 | 1 |
| AP3 | 5 | 0 | 5 | 4 | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| AP4 | 4 | 4 | 3 | 3 | 1 | 1 | 0 | 1 | 0 | 0 | 0 | 1 |
| AP5 | 5 | 2 | 5 | 3 | 2 | 0 | 1 | 0 | 1 | 0 | 0 | 0 |
| AP6 | 6 | 3 | 5 | 3 | 2 | 0 | 0 | 1 | 1 | 0 | 0 | 1 |
| GP1 | 5 | 0 | 4 | 3 | 2 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
| GP2 | 5 | 1 | 3 | 3 | 5 | 2 | 6 | 0 | 0 | 0 | 0 | 0 |
| GP3 | 6 | 3 | 2 | 2 | 4 | 2 | 4 | 0 | 0 | 0 | 0 | 0 |
| HP1 | 6 | 5 | 4 | 0 | 1 | 0 | 0 | 5 | 1 | 1 | 0 | 0 |
| MW1 | 6 | 0 | 5 | 0 | 5 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| MW2 | 6 | 4 | 5 | 1 | 3 | 3 | 2 | 0 | 0 | 0 | 0 | 0 |
| MW3 | 5 | 5 | 1 | 1 | 2 | 3 | 1 | 0 | 1 | 5 | 0 | 0 |
| MW4 | 6 | 4 | 3 | 4 | 0 | 6 | 1 | 3 | 0 | 0 | 0 | 0 |
| MW5 | 6 | 4 | 3 | 3 | 2 | 0 | 4 | 3 | 0 | 0 | 1 | 4 |
| MW6 | 6 | 4 | 6 | 3 | 1 | 2 | 4 | 2 | 0 | 0 | 0 | 2 |
| MW7 | 6 | 3 | 6 | 0 | 2 | 1 | 0 | 2 | 4 | 0 | 0 | 0 |
| RE1 | 5 | 3 | 4 | 0 | 0 | 1 | 0 | 0 | 2 | 5 | 0 | 0 |
| RE2 | 2 | 1 | 2 | 0 | 1 | 1 | 0 | 0 | 1 | 3 | 0 | 0 |
| RE3 | 5 | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 5 | 0 | 0 |
| GL1 | 6 | 1 | 3 | 0 | 5 | 2 | 2 | 0 | 0 | 0 | 0 | 1 |
| GL2 | 6 | 0 | 6 | 3 | 6 | 1 | 4 | 0 | 1 | 0 | 0 | 0 |
| GL3 | 6 | 2 | 4 | 4 | 4 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| GL4 | 5 | 2 | 1 | 0 | 3 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| GL5 | 6 | 1 | 1 | 3 | 6 | 3 | 4 | 0 | 3 | 0 | 0 | 0 |
| WR1 | 6 | 3 | 4 | 5 | 2 | 2 | 2 | 0 | 0 | 0 | 0 | 2 |
| WR2 | 5 | 2 | 0 | 2 | 3 | 0 | 0 | 1 | 0 | 0 | 0 | 2 |
| WR3 | 6 | 3 | 2 | 5 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 4 |

